

The potential of naturally occurring bacteria for the bioremediation of toxic metals pollution

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Abstract. An increase in industrialization and various kind of human activities added a huge amount of toxic heavy metals in the soil. As a result, toxic heavy metals in the environment may be adversely affects human being and aquatic ecosystem. Thus, it is very essential to understand mechanism of bioremediation through eco-friendly agent i.e. bacteria. Accumulation of high metal concentrations in soil above threshold limit causes lethal to bacterial communities in the environment. Few bacteria develop resistance mechanism to tolerate these toxic heavy metals and contain various methods to respond the metal stress. The present review emphasizes to understand the mechanism of bacterial resistance against toxic metals. Moreover, mechanism of bioaugmentation, biosorption, and bioaccumulation methods also described clearly.

Keywords: Bioremediation; Xenobiotic compounds; Heavy metals; Bacteria; Tolerance; PAHs.

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Introduction

Increase of agriculture and manufacturing industries have increased the emission of several vital xenobiotic compounds in the environment. Excess amount of toxic metal waste has decrease the level of drinking water in the environment and thus resulting in reduction of crop production

(Kamaludeen et al., 2003). Bioremediation exploits mainly bacteria to remediate the contaminated water and soil (Strong and Burgess, 2008). It relies on promoting the growth of bacterial consortia which is indigenous to the contaminated sites (Agarwal, 1998). Bacterial consortia can be develop through addition of nutrients by optimum temperature and moisture

conditions (Hess et al., 1997; Smith et al., 1998). Bacteria take contaminants as a nutrient source in bioremediation technology (Tang et al., 2007).

Bioremediation technology is used to remediate contaminated environment through bacterial consortia. Several scientists (Odu, 1978; Sloan, 1987; Ijah and Antai, 1988; Okpokwasili and Okorie, 1988; Barnhart and Meyers, 1989; Anon, 1990; Pritchard, 1991; Pritchard and Costa, 1991; Ijah, 2002, 2003) have revealed different applications of bacteria in the bioremediation of oil pollution with good results. Use of shellfish polluted by polycyclic aromatic hydrocarbons could be the reason of lung cancer in human beings (Law and Klungsoyr, 2000; Gaspare et al., 2009). Water contaminated through petroleum oil and hydrocarbons is very common in developed countries and it is very harmful to human health and aquatic ecosystem (Al-Baldawi et al., 2015). Biological bioremediation is an eco-friendly process which is very essential to remediate polluted aquatic ecosystem (Head and Swannell, 1999; Head et al., 2006; Beolchini et al., 2010; Dell'Anno et al., 2012). Furthermore, several authors demonstrated that petroleum pollution is very toxic to the environment (Lyons et al., 1999; Janjua et al., 2006). But fortunately, bacteria use these contaminants as a nutrient source (Atlas 1995; Oliveira et al., 2012). Many researchers demonstrated that both protozoa and algae are not responsible to degrade hydrocarbons (O'Brien and Dixon, 1976; Bossert and Bartha, 1984; Saadoun and Al-Ghzawi, 2005). Hassan et al. (2016) reviewed the various processes involved in bioremediation coupled with electrokinetics. Currently, Igris et al. (2018) demonstrated the bioremediation of heavy metals contaminated ecosystem from tannery wastewater. Saranya et al. (2017) also screened *Vibrio fluvialis* from industrial effluents for the bioremediation of mercury.

Background of heavy metals

Heavy metal usually occurs in its two forms i.e. bio-available and non-bioavailable forms. Their mobility depends on positive and negative charged component of salt (Sposito 2000). Cation exchange capacity, clay minerals, redox potential, buffering capacity, pH, organic matter, water content, temperature and bacterial activities determines the bio-availability of heavy metals in soil (Khan et al. 2009; Brown et al. 1999). Metals occur in soluble cationic forms under aerobic conditions, while on other side, found in carbonate or sulphide precipitates during anaerobic conditions. Khan et al. (2009) and Brown et al. (1999) studied the bio-availability of few heavy metals in following order: Zinc > Copper > Cadmium > Nickel. Nevertheless, the concentration of heavy metals varies drastically in the ecosystem. Therefore, these secondary pollutants get accumulate into food chain and finally into human food. Contamination of agriculture soil with heavy metals has become a major problem for the sustainability of the environment. Thus, assessment of heavy metal bio-availability enables to evaluate the impact of metals on soil bacteria and it demonstrate the use of bioremediation technologies to clean heavy metals pollution from soil. Lead and Cadmium are main two heavy metals which are the most toxic because they have no role in the metabolism of living beings. Both said metals are distributed hugely in the environment so it is very important to remediate the environment from these two metals.

Lead

Lead (Pb) is very persistent kind of secondary pollutant in environment. Human being is exposed to Pb through paint, cosmetics, folk remedies, lead batteries and other food supplements etc. ATSDR (2007) reported that inorganic Pb usually absorb by oral and dermal exposure through the respiratory

system. Absorption of lead acetate and lead chloride was observed higher in children than in adults (ATSDR, 2007). According to ATSDR (2007) in adults, around 92% of body burden of Pb remains confined to bones as compared to 77% in children. Pb is a ubiquitous secondary pollutant which is well known to change the hematological system by inhibiting the function of several enzymes involved in the biosynthesis of haem (Hernberg and Nikkanen, 1970; Millar et al., 1970). Pb inhibits the entry of calcium ions (Ca^{2+}) into cells with synaptogenesis. In blood, Pb is primarily found in red blood cells (RBC) and also responsible for the inhibition of human neurotransmitter system. Pb destabilizes the cellular membrane of RBC and reduces the fluidity of cell membrane and as a result rate of erythrocyte hemolysis increases which leads to anemia (Bellinger and Bellinger, 2006; Needleman, 2004). Pb also mimics to Ca^{2+} and disturb calcium homeostasis which leads to the production of neurotoxin having adverse effects on central nervous system. Pathogenesis of Pb toxicity is multifactorial and it induces oxidative stress by generation of reactive oxygen species (ROS). It reduces antioxidant defense mechanism of cells by interrupting with essential metals needed for enzyme activities and also alters membrane integrity and fatty acid composition (Gurer and Ercal 2000). Metabolism of inorganic Pb composed of formation of complexes with protein and non-protein ligands. Inorganic Pb compounds are actively metabolized in liver through oxidative dealkylation by P450 enzymes.

Cadmium

Cadmium (Cd) occurs naturally in the earth's crust in association with copper, zinc and lead ores. It is generally found in cadmium-nickel battery manufacture, nonferrous metal, waste incineration, refining, phosphate fertilizers and disposal of environment. Cadmium expose mainly through

cigarette smoke, chocolate, mushroom and seafood in human beings (EFSA, 2009). Excess accumulation of Cd salt causes severe problems like osteoporosis, anemia and renal tubular injury (Jarup, 2003). Miura (2009) revealed that Cd is a potent carcinogen which associated with cancer of kidney, pancreas and lung and was classified by International Agency for Research on cancer. Still, molecular mechanism of Cd based carcinogenesis is not yet understood. Replacement of zinc in zinc finger structures was planned to understand the mechanism of mutagenic effect of Cd salt.

Mechanism of bacterial resistance against heavy metal stress

Accumulations of toxic heavy metals in the soil and their absorption by the plants have become very important concern for environmental scientist. Unlike other secondary pollutants, toxic heavy metals could also biodegrade to less toxic products (Kumar et al. 2011). Some heavy metals e.g. nickel, zinc, copper and chromium are essential micronutrients which require for the growth of plants, animals and microorganisms (Olson et al. 2001). While, cadmium, lead and mercury have no biological or physiological role in living beings (Gadd 1992). Therefore, higher concentration of cadmium, mercury and lead have very adverse effects on the bacterial community in soil by three different ways, (1) It leads to reduction of total bacterial biomass (Giller et al. 1998) (2) It reduces the population of specific bacteria (Chaudri et al. 1993) (3) It alters the structure of bacterial community (Gray and Smith 2005). Hence, at higher concentration of toxic metal ions may be inhibit the growth of bacterial population through interruption of normal activities as shown in figure 1. But, few potential bacteria can develop resistance or tolerance against high concentration of metals. Generally, resistance may be defined as ability to deal with pollutant

toxicity with help of intrinsic properties of bacteria. While, tolerance is the phenomena in which bacteria get survive even in higher concentrations of heavy metals by the mechanism of detoxification (Ahemad et al. 2009). Therefore, toxic heavy metals needed to either removed completely or transformed to be less toxic forms. Bacteria develop several mechanisms to acclimatize under metal stressed conditions to tolerate the uptake of toxic metal ions (Nies 1999). Five different mechanisms include following processes;

(1) Accumulation-Bacteria forms complex with metal binding proteins (e.g. metallothionins, a low molecular wt. proteins) (Kao et al. 2006; Umrana et al. 2006), (2) Exclusion-Toxic metal ions are removed from target sites, (3) Biotransformation-Toxic metal ions are allowed to convert into toxic forms, (4) Methylation and Demethylation-One or more of these defence mechanisms allows bacteria to function metabolically in metals polluted environment, (5) Extrusion-Metals are pushed out of the cell by chromosomal mediated events.

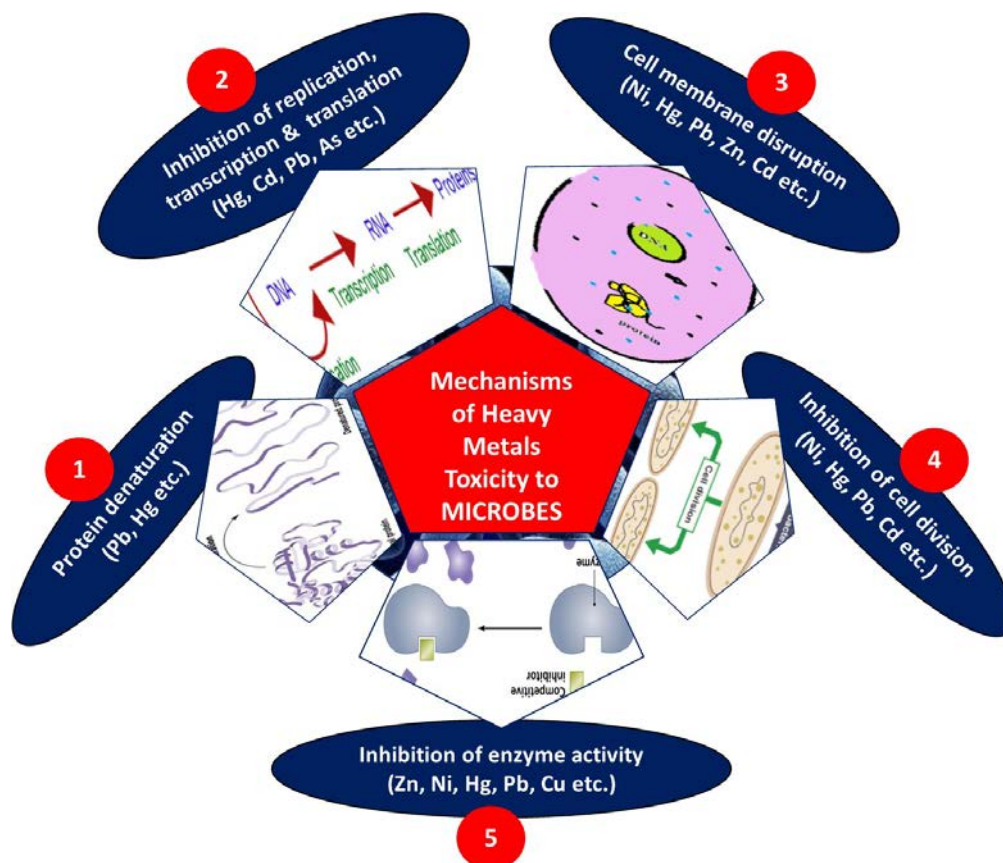


Figure 1. Mechanisms of Heavy metal-toxicity to microorganism.

Bacterial resistance mechanism is usually present on plasmid and chromosome part of cell. Bacteria acclimatize their resistance against heavy metals through gene transfer or spontaneous mutation. Nies (1999)

demonstrated CZC metal tolerance mechanism in gram negative bacteria i.e. *Ralstonia eutropha* and CZC genes are responsible for a cation-proton antiporter (czcABC), which excludes zinc, cadmium and cobalt metals. A similar

kind of mechanism was reported in *Alcaligenes xylooxidans* which provides resistance against nickel, cobalt and cadmium. Furthermore, Roane and Pepper (2002) reported that cadmium resistance mechanism is different in gram positive bacteria (e.g. *Staphylococcus*, *Bacillus* or *Listeria*) and it was through Cd efflux ATPase. They demonstrated plasmid encoded resistance mechanism through ATPase,

chemiosmotic ion or proton pumps in chromium, arsenic and cadmium in some bacteria. Lloyd and Lovley (2001) studied unique properties of bacterial resistance for the remediation of toxic metal contaminated sites. Therefore, bacteria can interact among heavy metals through several mechanisms. There is a list of bacteria which are able to accumulate heavy metals toxicity (Table 1).

Table 1. List of some comprehensively investigated heavy metal accumulating bacteria. Metals, bacteria and their source, MRL/uptake efficiency and removal percentage with removing time has been shown.

Metal	Microorganism	Source	MRL/uptake efficiency	Removal (%)	References
Cd	<i>Bacillus</i> strain H9	Metal-contaminated soil	284 µg/mL	35 (50 h)	Roane et al., 2001
	<i>Aspergillus terreus</i>	Industrially polluted sediments	124 ppm/g	72 (14 days)	Massaccesi et al., 2002
Cr	<i>Pseudomonas aeruginosa</i>	Wastewater	0.079 mg/mL	46 (2 days)	Hassen et al., 1998
	<i>Aspergillus niger</i>	Acquired	5.2-6.5 mg/g	22-35 (7 days)	Dursun et al., 2003
Pb	<i>Pseudomonas aeruginosa</i> PU21 (Rip64)	-	0.6 mg/mL (112 mg/g)	82 (3 days)	Chang et al., 1997
	<i>Aspergillus niger</i>	Acquired	5.2-33.4 mg/g	14-86 (7 days)	Dursun et al., 2003
Cu	<i>Thiobacillus ferrooxidans</i>	Adopted	0.12 mg/mL (704 mg/g)	24 (18 min)	Boyer et al., 1998
	<i>Schizosaccharomyces pombe</i>	Acquired	0.5-1.2 mg/g	12-24 (5 days)	Donmez and Aksu, 1999
Ni	<i>Pseudomonas</i> spp.	Local isolate	72.9 mg/g	97 (5 days)	Magyarosy et al., 2002
	<i>Candida</i> spp.	Sewage	10.2-42.4 mg/g	28-58 (5-12 days)	Donmez and Aksu, 2001

Abbreviations: MRL: Maximum Resistance Level.

Classification of bioremediation

Bioremediation can be classified to different methods such as biosorption, biomineralization, bioaccumulation, bioleaching and biotransformation. Bacteria use the heavy metals from soil as nutrients source for their growth and development. They have capability to reduce or oxidize transition metals. Bioremediation can be made successful by understanding the mechanism of bacterial growth in the contaminated sites. Sikkema et al. (1995) revealed that bacterial cells develop defense mechanisms against contaminants by forming outer cell membrane protective material.

Biosorption

Heavy metals form interactions with the binding sites present in the cellular structure of bacteria through biosorption method. Extracellular polymeric substances (EPS) of bacterial cell walls have significant properties for metal adsorption and EPS revealed a great ability to form complex through specific mechanisms for the precipitation of metals (Guine et al., 2006; Comte et al., 2008; Fang et al., 2010). Fang et al. (2011) characterized and quantified proton and adsorbed on bacterial cells and studied the importance of EPS molecules to remove the toxic metal. Therefore, it is very

important to understand genomic characteristic and metabolic pathway of bacteria used in metal adsorption (Kinya and Kimberly, 1996; Carter et al., 2006; Gan et al., 2009; Haritash et al., 2009; Onwubuya et al., 2009).

Metal precipitation

Several bacteria can be used to immobilize toxic heavy metals through the process of their lower redox state and producing less bioactive metal species. Metal precipitation is a very common activity which is the result of reduction of metabolic processes unrelated to the transformed metals. Roane and Pepper (2000) demonstrated that Cr(VI) get reduces to insoluble Cr(III) through bacterial respiration or by indirect reduction of sulphide. Metal precipitation strategy has become more interested for environmental scientist by the formation of metal sulphides and ferrous through indirect reduction. However, dissimilatory metal reduction can also be safely used for decontamination.

Bioremediation by physio-bio-chemical mechanism

Biosorption process involves higher affinity of a biosorbent for metal ions, continued until equilibrium is established between two components (Das et al. 2008). Later Chen and Wang (2007) and Talos et al. (2009) revealed that *Saccharomyces cerevisiae* acts as a biosorbent for the removal of Cd and Zn ions through ion exchange mechanism. *Cunninghamella elegans* emerged as a promising sorbent against heavy metals released through textile waste water (Tigini et al., 2010). Both active and passive modes of toxic metal bioremediation may be called bioaccumulation (Brierley, 1990). Moreover, Pinedo et al. (2009) revealed that fungi have potential to act as biocatalysts to access heavy metals and convert them into less toxic form. Some fungi such as *Botryosphaeria rhodina*, *Pleurotus pulmonarius* and *Allescheriella*

species have metal binding capacity (D'Annibale et al., 2007). Lead contaminated soils can be biodegraded by fungal species e.g. *Aspergillus parasitica* and *Cephalosporium aphidicola* through biosorption process (Tunali et al., 2006; Akar et al., 2007). Hg tolerant fungi (*Neocosmospora vasinfecta* and *Verticillium terrestris*) were able to biotransform Hg²⁺ state to its less toxic form (Kelly et al., 2006). Generally, contaminants are hydrophobic in nature and these materials appear to be taken up by bacteria through the secretion of few biosurfactants. Biosurfactants forms stronger ionic bonds with heavy metals and form complexes due to low interfacial tension (Thavasi, 2011). Bioremediation involves aerobic or anaerobic microbial activities. Aerobic degradation adds oxygen into the reactions mediated through different enzymes like hydroxylases, oxidative dehalogenases and chemically reactive oxygen. Anaerobic degradation of pollutants involves initial activation reactions through oxidative catabolism which is mediated by anoxic electrons. Immobilization is the process in which mobilization of toxic metals get reduce from polluted sites through physical or chemical state. Solidification treatment could be used to mixing of chemical agents at pollutants sites (Evanko and Dzombak, 1997). Bacteria get mobilize heavy metals from contaminated sites through chelation, redox transformation and methylation of toxic metals. Furthermore, Garbisu et al. (2001) revealed that heavy metals can't be destroyed completely, but metals could be converted to its precipitated and less toxic forms. Bacteria remove toxic metal ions by using the mechanism which utilizes to derive energy from redox reactions in enzymatic and non-enzymatic processes. Silver (1996) studied two mechanisms which are responsible for the development of resistance in bacteria: 1. Detoxification is the process in which toxic metal gets transform to less toxic form. 2. Active

efflux pumping of toxic metal ions from cell. Bacteria act as an oxidizing agent for heavy metals and lose electrons, which are accepted by alternative electron acceptors e.g. nitrate, sulphate and ferric oxides. Oxygen acts as an electron acceptor in aerobic conditions, while it oxidizes inorganic contaminants in anaerobic conditions. Bacteria take energy for their growth through oxidizing organic compound from Mn(IV) or Fe(III) as an electron acceptor (Lovley and Phillips, 1988). Anaerobic degradation of organic pollutant is stimulated with higher availability of iron for bacterial reduction (Spormann and Widdel, 2000). Lovely (2002) demonstrated that metals are used as terminal electron acceptors and known as dissimilatory metal reduction.

Bacteria convert the state of metals and reduce the uranium phase from U⁶⁺ to U⁴⁺ (Lovley et al., 1991). Different defense mechanisms viz., exclusion, formation of binding proteins and complex formation decrease the effect of stress produced from toxic heavy metals (Gómez Jiménez-T et al., 2011). Heavy metal accumulation has been studied through expression of metal binding protein and peptides in bacteria (Cobbett and Goldsbrough, 2002). The expression of *smtA* gene and production of metal binding protein was studied in

Synechococcus sp. (Huckle et al., 1993). Later, *Ralstonia eutropha* was genetically manipulated to express mouse metallothionein protein on the surface of cell and reduce the toxic effect of the Cd²⁺ from the contaminated site (Valls et al., 2000). Mejare and Bulow (2001) studied the expression of different proteins and peptides to regulate the range of accumulation of cadmium in *Escherichia coli*. Natural tolerance pathways for metal toxicity have been regulated by metalloregulatory protein in bacteria (Singh et al., 2008).

Genetic Engineering for bioremediation processes

Genetically modified microbes are organisms in which genetic material has been manipulated using recombinant DNA technology (RDT) to identify competent strain to remediate hazardous contaminants present in ecosystem (Sayler and Ripp, 2000). The genetically engineered potential bacteria which remove the toxicity of heavy metals are given in Table 2. Nowadays, genetic engineering has been developed different bacterial biosensors to measure the level of contamination in environment. Several biosensors have been made to estimate heavy metal concentration like cadmium, mercury, arsenic and copper (Verma and Singh, 2005; Bruschi and Goulhen, 2006).

Table 2 Genetically engineered bacteria for bioremediation of heavy metals.

Heavy metal	Initial conc. (ppm)	Removal efficiency (%)	Genetically engineered bacteria	Expressed gene	References
As	0.05	100	<i>E. coli</i> strain	Metalloregulatory protein ArsR	Kostal et al. 2004
Cd ²⁺	-	-	<i>E. coli</i> strain	SpPCS	Kang et al. 2007
Cr ⁶⁺	1.4–1000	100	<i>Methylococcus capsulatus</i>	CrR	Hasin et al. 2010
Cr	-	-	<i>P. putida</i> strain	Chromate reductase (ChrR)	Ackerley et al., 2004
Cd ²⁺ , Hg	-	-	<i>Ralstonia eutropha</i> CH34, <i>Deinococcus radiodurans</i>	merA	Valls et al. 2000
Hg	-	-	<i>E. coli</i> strain	Organomercurial lyase	Murtaza et al., 2002
Hg	7.4	96	<i>E. coli</i> JM109	Hg ²⁺ transporter	Zhao et al., 2005
Hg	-	-	<i>Pseudomonas</i> K-62	Organomercurial lyase	Kiyono and Pan, 2006
Hg	-	-	<i>Achromobacter</i> sp A022	mer	Ng et al. 2009
Ni	145	80	<i>P. fluorescens</i> 4F39	Phytochelatin synthase (PCS)	Lopez et al. 2002

Bioremediation by bio-augmentation

Several authors demonstrated that addition of nutrients is essential to

enhance bacterial growth and biodegradation activity. Wu et al. (2012) revealed the surface activity of salt tolerant *Serratia Spp.* and crude oil

degradation in saline soil. The novel strain of *Serratia Spp* BF40 was isolated from crude oil contaminated soils and evaluated for its surface activity, salt tolerance from saline soil. The authors also suggested that strain reduce the surface of oily soil surface and concluded that using bacterial strain with bio-surfactant producing capability was efficient. *Candida tropicalis* (SK 21 strain) was used for bio-augmentation ability by Fan et al. (2013). Genetic engineering of rhizospheric bacteria with plant associated degradation of contaminants has become new technology to remediate metal contaminated sites (Divya and Kumar, 2011). *Escherichia coli* and *Moreaxella* sp. expressed phytochelatin 20 on cell surface to accumulate Hg and Cd (Bae et al., 2001, 2003). Several scientists tried to understand molecular approaches in *Escherichia coli*, *Bacillus subtilis* and *Pseudomonas putida*.

Bio-stimulation by using inorganic nutrients

Inorganic fertilizers have been utilized in the form of bio-stimulation agents globally. Chorom et al. (2010) demonstrated the efficiency of inorganic fertilizer to enhance microbial degradation of petroleum hydrocarbons in contaminated soil. Agarry and Ogunleye (2011) used inorganic fertilizers as independent bio-stimulation variables. Venosa et al. (2002) reported that bacteria are easily degraded light end hydrocarbon than heavy end hydrocarbons. Thus, addition of inorganic fertilizers was more effective in the improvement of biological degrading activity.

Conclusions

Bioremediation is a very important tool used to clean up the contaminants present in the environment. The initiation of bioremediation has been started for several years. However, some applications are relatively new and many

other applications are emerging or being developed. Bioremediation occurs when micro-organisms can biodegrade the secondary pollutant like heavy metals, polycyclic aromatic hydrocarbons and pesticides. This process may be aerobic or anaerobic that depends on the availability of micro-organisms and electron acceptors. These methods of bioremediation may be natural or improved by engineered bioremediation. This technology is very efficient and cost effective to treat contaminated water and soil. This review concluded that organic and inorganic nutrients could promote bacterial growth and degradation of toxic metal pollution in the environment. We also elucidated the fact that biodegradation of toxic heavy metals in the soils is feasible with bacterial metabolism.

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Conflict of interest

The authors declare that there is no conflict of interest.

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